

Feasibility of seed bank for restoration of salt marsh: a case study around the Gwangyang Bay, southern Korea

Seonmi Lee¹, Yong-Chan Cho² and Chang Seok Lee^{3,*}

¹Department of Biology, Graduate School of Seoul Women's University, Seoul 139-774, Korea

²Department of Forest Resource Conservation, Korea National Arboretum, Pocheon 487-821, Korea

³Department of Life and Environmental Engineering, Seoul Women's University, Seoul 139-774, Korea

Abstract

Salt marsh is an important transitional zone among terrestrial, riverine, and marine ecosystems and is a productive habitat that interacts extensively with adjacent landscape elements of estuarine and coastal ecosystems. Nowadays, in addition to various human activities, a variety of natural processes induce changes in salt marshes. This study aims to provide background information to restore disturbed salt marshes and to propose their ecological restoration using seed banks. The study area is a prepared area for the Gwangyang Container Port located in the southern Korea. This area was formed by accumulating mud soils dredged from the bottom of the forward sea. This land was created in a serial process of preparing the Gwangyang container port and the salt marsh was passively restored by seeds buried in mud soil dredged from seabed. As a result of stand ordination based on vegetation data collected from the land, stands were arranged according to tolerance to salinity in the order of *Suaeda maritima*, *Salicornia europaea*, and *Phragmites communis* communities on the Axis 1. Landscape structure of the projected area was analyzed as well. Edges of the projected area were divided from the marginal waterway by the dike. Four types of vegetation appeared on the dike: *Alnus firma* plantation, *Robinia pseudoacacia* plantation, *Lespedeza cyrtobotrya* plantation, and grassland. In the more internal areas, two types of vegetation sequences appeared: *Aster tripolium* community-*Suaeda glauca* community-*Salicornia europaea* community sequence and *Aster tripolium* community-*Suaeda maritima* community-*S. europaea* community sequence. Mixed community showed the highest species diversity ($H' = 0.86$) and *S. europaea* community showed the lowest ($H' = 0.0$). Evenness is the highest in Mixed community ($J' = 2.26$) and the lowest in *S. maritima*-*S. europaea* community ($J' = 0.0$). Several plant communities were successfully established on the land created by mud soil dredged from the bottom of Gwangyang Bay. Moreover, community diversity in this area approached a similar level with those from other studies involving natural salt marshes. Therefore, restoration effect based on community diversity obtained in our study can be evaluated as a successful achievement. In this respect, although most salt marshes in Korea and other places worldwide have been destroyed or disturbed by excessive land use, feasibility of seed bank as a restoration tool is greatly expected.

Key words: Gwangyang bay, halophyte, restoration, salt marsh, seed bank

INTRODUCTION

Salt marshes appear along the edges of shallow seas with soft sediment (Eisma et al. 1998) and are important transitional zones among terrestrial, riverine, and marine

ecosystems. Salt marshes include distinctive natural resources and maintain key ecosystem processes (Wall et al. 2001). They develop on coastal shores where they are

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***Corresponding Author**

E-mail: leecs@swu.ac.kr

Tel: +82-2-970-5666

www.kci.go.kr

sufficiently protected from wave and tidal energy to allow sediment suspended in the water column to settle out and accumulate (Chung et al. 2004, Hughes and Paramor 2004, Nottage and Robertson 2005).

Salt marshes develop well in the western and southern coasts of Korea because of the big difference between the flow and ebb tides thereby occupying 2.7% of total national area (Han 2008). Salt marshes are ecologically important and productive habitats that interact extensively with adjacent landscape elements of estuarine and coastal ecosystems. The most obvious biological characteristic is the unique plant communities including many species, which are found only in tidally influenced areas (Nottage and Robertson 2005). Nowadays, however, salt marshes are changing by a variety of natural processes and threatened by human activities (Doody 2001, Nottage and Robertson 2005). In most areas of Korea, salt marshes experience human exploitation such as salterns, farms, reclamations, and so on. A lot of salt marshes have disappeared already but the damage is in progress because of the high land use intensity (Ministry of Environment 2006).

Ecological restoration is the process of assisting recovery of degraded, damaged or destroyed ecosystems (Society for Ecological Restoration International Science & Policy Working Group 2004) and the return to a similar approximation of their original state (National Research Council 1992). The goal of this is to reestablish a complete functional ecosystem (Stanturf et al. 2001).

Seed banks are a group of non-germinated but viable seeds stored in the soil. They can remain dormant for many years to decades until surrounding environments are appropriate for germination (Grime 1989, Simpson et al. 1989). These dormant seed banks can have ecological and evolutionary influences on actual plants above ground (Kalisz 1990). Therefore, they are recognized as a critical element in management plans of a given ecosystem (Cox and Allen 2008). In addition, after destruction or disturbance of an ecosystem, they play an important role in natural regeneration (Grime 1981, Roberts 1981). The value of the seed banks for restoration is different according to the type and length of disturbance of the site (Middleton 1999).

To date in Korea, most studies on salt marshes have focused on the environmental factors affecting vegetation (Lee and Kim 1988, Lee et al. 2009) such as characteristics and distribution of halophyte (Ihm 1989, Lee 1989, Ihm et al. 1995a, 1995b, 1998a, 1998b, Kim et al. 2005, Han 2008). Therefore, many studies haven't been done regarding salt marsh restoration, particularly through the use of seed

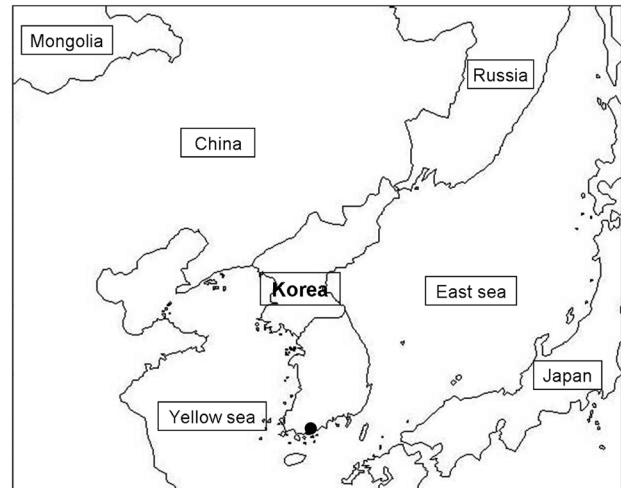


Fig. 1. A map showing the study area.

banks.

This study aims to diagnose feasibility of ecological restoration of artificially disturbed coastal areas. Further, this study aims to evaluate restoration effects based on naturally established species composition and diversity of vegetation in a salt marsh created by seeds buried in the mud soil dredged from the seabed.

MATERIALS AND METHODS

Study area

The study area is located around the Gwangyang Container Port, Southern Korea (latitude N 34°54'-N 35°10' and longitude E 127°31'-E 127°38'). The land of this study area was created by accumulating mud soil dredged from the seabed to ensure water depth in the Gwangyang container port (Fig. 1). This land was the bare flat in the early days, but it didn't remain for a long time because of the abundant seeds that were buried in the dredged soil. This study was carried out in the seventh year since the flat was created. The study area at the time of the survey was covered with salt marsh vegetation and thus resembled a typical salt marsh.

This area is surrounded by Gwangyang-si, Suncheon-si, Yeosu-si, Namhae-county, and Hadong-county and belongs to the warm-temperate forest region (Yim and Kira 1975). Plant communities dominating in this region are *Pinus thunbergii* community, *P. densiflora* community, *Quercus variabilis* community, *Q. acutissima* community, *Q. serrata* community, and so on (Ministry of Environ-



Fig. 2. A physiognomic vegetation map showing the projected area for the Gwangyang Container Port.

ment 2006). Mean annual precipitation and temperature at Suncheon weather station, which is the closest weather station to the study area, were 1,487 mm and 12.5°C, respectively.

Methods

Creation of salt marsh

The land of salt marsh of our interest was created by accumulating mud soil dredged from the bottom of the Gwangyang Bay to ensure water depth in the Gwangyang container port. Salt marsh vegetation was restored passively by seeds buried in the mud soil dredged from the seabed.

Vegetation sampling and data analysis

Vegetation survey was carried out in the salt marsh restored passively by seeds buried in the mud soil dredged from the seabed. Vegetation map was prepared by interpretation of IKONOS satellite image and color aerial photograph, and field check. The vegetation map was constructed with ArcView GIS (Environmental System Research Institute 1996).

Vegetation sampling was conducted from May to September, 2007 on 65 plots. Vegetation survey was carried out by recording the cover class of plant species that ap-

peared in randomly installed 1 m × 1 m quadrats. Cover data were transformed from the ordinal scale of Braun-Blanquet (1964), and subjected to ordination. Differences in species composition were analyzed with detrended correspondence analysis (Hill 1979) by using PC-Ord 4 program (McCune and Mefford 1999). Species diversity was compared by Shannon-Wiener (Magurran 2004) and evenness (Pielou 1966) indices.

RESULTS

Landscape analysis

Landscape structure of the projected area and its surrounding areas are depicted in Fig. 2. The edges of the projected area were divided from the marginal waterway by the dike. Vegetation was introduced on the dike and formed a hedgerow. This hedgerow was constructed around the projected area except the southern road side. Four types of vegetation appeared on the dike: *Alnus firma* plantation, *Robinia pseudoacacia* plantation, *Lespedeza cyrtobotrya* plantation, and grassland. Dominated vegetation was established bordering the hedgerow. *Phragmites communis* community occupied the adjacent areas of the living hedgerow in a gentle slope. In the

more internal areas, the following two types of vegetation sequences appeared from the marginal toward the central parts: *Aster tripolium* community-*Suaeda glauca* community-*Salicornia europaea* community sequence

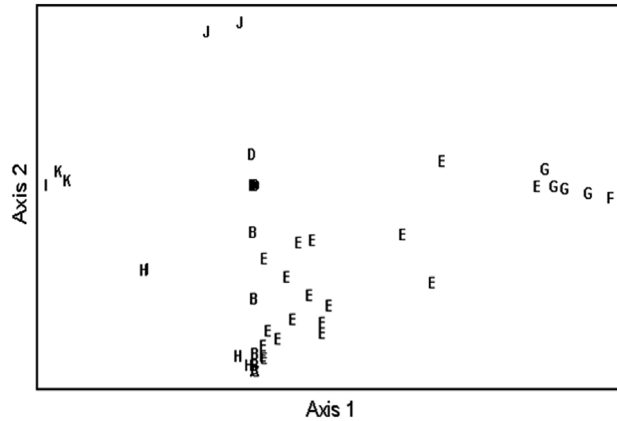


Fig. 3. Detrended correspondence analysis ordination based on the vegetation data collected in the 65 stands. A, *Salicornia europaea* community; B, *Salicornia europaea*-*Aster tripolium* community; C, *Salicornia europaea*-*Suaeda glauca* community; D, *Aster tripolium*-*Salicornia europaea* community; E, mixed community; F, *Phragmites communis* community; G, *Phragmites communis*-*Suaeda glauca* community; H, *Salicornia europaea*-*Suaeda maritima* community; I, *Suaeda maritima*-*Salicornia europaea* community; J, *Aster tripolium*-*Suaeda maritima* community; and K, *Suaeda maritima*-*Phragmites communis* community.

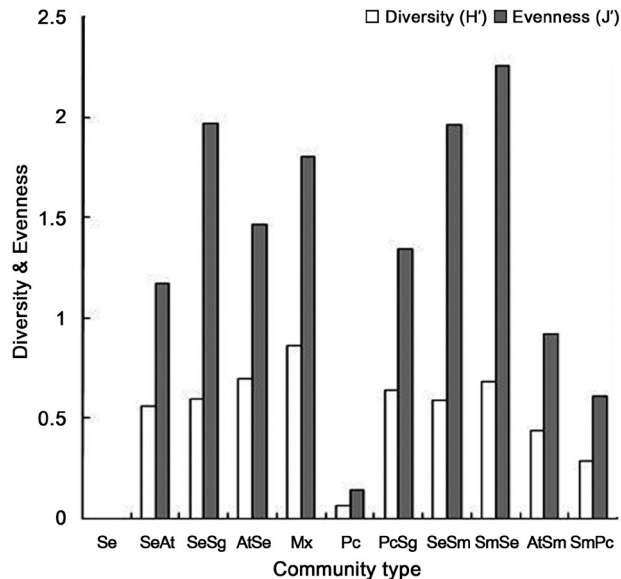


Fig. 4. Shannon-Wiener (H') and evenness (J') indices of the salt marshes vegetation. Se, *Salicornia europaea* community; SeAt, *S. europaea*-*Aster tripolium* community; SeSg, *S. europaea*-*Suaeda glauca* community; AtSe, *Aster tripolium*-*S. europaea* community; Mx, mixed community; Pc, *Phragmites communis* community; PcSg, *P. communis*-*S. glauca* community; SeSm, *S. europaea*-*Suaeda maritima* community; SmSe, *S. maritima*-*S. europaea* community; AtSm, *A. tripolium*-*S. maritima* community; SmPc, *S. maritima*-*P. communis* community.

and *Aster tripolium* community-*Suaeda maritima* community-*Salicornia europaea* community sequence. The former appeared in the west of the projected area, which was a little dry and the latter appeared in the east of that, which was relatively wet. On the other hand, *Triglochin maritimum* is growing in the pool remaining in a part of the study area but was not expressed in a vegetation map because the area is too small.

P. thunbergii community, *P. densiflora* community, and *P. rigida* plantation appeared in the surrounding forested area. In addition, cultivated land, orchard, and residential area existed in the lower forests and plains.

Species composition

Stand ordination based on vegetation data was carried out in order to compare species composition among plant communities established in artificially constructed salt marshes around the Gwangyang Container Port (Fig. 3). The eigen values of axes 1 and 2 were 0.813 and 0.386, respectively. According to tolerance to salinity, stands were arranged in the order of *Suaeda maritima*, *Salicornia europaea*, and *Phragmites communis* communities on the Axis 1. Significant relations were not found with the Axis 2.

Species diversity and evenness

Shannon-Wiener and evenness indices were compared among plant communities (Fig. 4). Mixed community showed the highest species diversity ($H' = 0.86$) and *S. europaea* community showed the lowest ($H' = 0.0$). The Shannon-Wiener index of all communities was below 1.0 because each community was composed of only a few species adapted to the characteristic environment of salt marshes. Evenness was the highest in Mixed community ($J' = 2.26$) and the lowest in *S. maritima*-*S. europaea* community ($J' = 0.0$).

DISCUSSION

Comparison of vegetation types

In the previous studies on salt marshes in Korea, Ihm (1989), Lee (1989), Ihm et al. (1998a), and Han (2008) classified fourteen, eight, nine, and six halophyte communities, respectively. Plant communities that they reported were *S. europaea* community, *Suaeda japonica* community, *S. maritima* community, *S. glauca* community, *Zoysia*

sinica community, *Artemisia scoparia* community, *Limonium tetragonum* community, *A. tripolium* community, *Atriplex gmelinii* community, *T. maritimum* community, *Carex scabrifolia* community, *P. communis* community, *Phacelurus latifolius* community, *Scirpus maritimus* community, *Typha orientalis* community, *Conyza bonariensis* community, and *Artemisia fukudo* community. These plant communities were usually expressed as pure stands. In our study area, five kinds of halophytes formed communities and those communities were usually mixed types rather than pure ones. A short period of seven years of restoration created several communities, which corresponded to 29.4% of total halophyte communities classified at salt marshes in Korea. Comparing the result with each study area mentioned above, community diversity in this area approached a similar level except for Ihm's (1989) result. However, it should be noted that Ihm's (1989) result was obtained from salt marshes throughout the entire South Korea. Considering the facts, restoration effect obtained in our study can be evaluated as a successful achievement.

Restoration of salt marshes using seed banks

Bradshaw (1984) suggested three methods for ecosystem restoration. First is a passive restoration such as natural recovery. Second is implementing minimum intervention to accelerate succession and third is active intervention such as sowing seeds, planting trees and using seed banks. In general, the success of salt marshes restoration depends on the three factors (Wolters and Bakker 2002), namely the seed banks (Willems and Bik 1998), the presence of species pool (Morton and Law 1997, Zobel 1997, Zobel et al. 1998), and the ability of species to disperse (Willson and Traveset 2000, White et al. 2004). It is believed that even if halophytes have disappeared, their seeds may be buried in the soil seed bank. Consequently, they can germinate when the surrounding environment becomes appropriate for growth (Fenner 1985). Recently, studies on active restoration using seed banks and the potential contribution of seed banks for restoration have been gradually increasing (Augusto et al. 2001, Richter and Stromberg 2005, Roovers et al. 2006, Koh 2007, Lee 2007, Zhan et al. 2007, Lee et al. 2008).

Salt marshes are one of the most disturbed places experiencing both natural and anthropogenic disturbances (Wang et al. 2009). Seed banks are very useful for recovery of annual plants after disturbances (Baskin and Baskin 1980, Lee et al. 2008) and for the design of restoration projects (Richter and Stromberg 2005). Dormant seeds

have significant ecological and evolutionary influences on plant population and community dynamics (Kalisz 1990). A few studies have mentioned that most seeds in salt marshes are transient and short longevity (Thompson et al. 1997, Wolters and Bakker 2002). However, several plant communities were established successfully from soils dredged from the bottom of Gwangyang Bay even though the original purpose of this work was not restoration. Moreover, community diversity is similar to the results of other studies. Since there is a great number of salt marshes in Korea and other places worldwide which have been destroyed or disturbed by excessive land use, the use of seed banks as a restoration tool is expected to increase. In addition, it would be applicable to restore landfill front in the reclaimed area such as Saemangeum. In fact, case studies executed in the United States and Netherlands showed that soil seed banks contributed significantly to re-establishment of salt-marsh communities (Turner et al. 1988, Toth 1996, Bakker et al. 2002).

LITERATURE CITED

- Augusto L, Dupouey JL, Picard JF, Ranger J. 2001. Potential contribution of the seed bank in coniferous plantations to the restoration of native deciduous forest vegetation. *Acta Oecol* 22: 87-98.
- Bakker JP, Esselink P, Dojkema KS, van Duin WE, de Jong DJ. 2002. Restoration of salt marshes in the Netherlands. *Hydrobiologia* 478: 29-51.
- Baskin JM, Baskin CC. 1980. Role of seed reserves in the persistence of a local population of *Sedum pulchellum*: a direct field observation. *Bull Torrey Bot Club* 107: 429-430.
- Bradshaw AD. 1984. Ecological principles and land reclamation practice. *Landsc Plann* 11: 35-48.
- Braun-Blanquet J. 1964. *Pflanzensoziologie*. 3rd ed. Springer, Wien.
- Chung CH, Zhuo RZ, Xu GW. 2004. Creation of *Spartina* plantations for reclaiming Dongtai, China, tidal flats and offshore sands. *Ecol Eng* 23: 135-150.
- Cox RD, Allen EB. 2008. Composition of soil seed banks in southern California coastal sage scrub and adjacent exotic grassland. *Plant Ecol* 198: 37-46.
- Doody JP. 2001. *Coastal Conservation and Management: An Ecological Perspective*. Kluwer Academic Publishers, Boston, MA.
- Eisma D, de Boer PL, Cadée GC. 1998. *Intertidal Deposit: River Mouths, Tidal Flats, and Coastal Lagoons*. CRC Press, Boca Raton, FL.
- Environmental System Research Institute (ESRI). 1996. Ar-

- cView GIS. Environmental System Research Institute, Inc., New York.
- Fenner M. 1985. Seed Ecology. Chapman & Hall, London.
- Grime JP. 1981. The role of seed dormancy in vegetation dynamics. *Ann Appl Biol* 98: 555-558.
- Grime JP. 1989. Seed banks in ecological perspective. In: Ecology of Soil Seed Banks (Leck MA, Parker VT, Simpson RL, eds). Academic Press, London, pp 15-22.
- Han YU. 2008. The characteristics of halophyte vegetation of salt marshes in the southern and western coasts of Korea. Ms Thesis. Mokpo National University, Muan, Korea.
- Hill MO. 1979. DECORANA: A FORTRAN Program for Detrended Correspondence Analysis and Reciprocal Averaging. Cornell University, Ithaca, NY.
- Hughes RG, Paramor OAL. 2004. On the loss of saltmarshes in south-east England and methods for their restoration. *J Appl Ecol* 41: 440-448.
- Ihm BS. 1989. Distribution of coastal plant communities in response to soil water potential and plant osmotic adjustment. PhD Dissertation. Seoul National University, Seoul, Korea. (in Korean with English abstract)
- Ihm BS, Lee JS, Kim HS, Kwak AK, Ihm HB. 1995a. Adjustment of three halophyte to changes of NaCl concentrations. *Bull Inst Litt Environ Mokpo Natil Univ* 12: 1-10.
- Ihm BS, Lee JS, Kim HS, Kwak AK, Ihm HB. 1995b. Distribution of coastal plant communities at the salt marshes of Mankyung and Donjin River estuary. *Bull Inst Litt Environ Mokpo Natl Univ* 12: 11-28.
- Ihm BS, Lee JS, Kim JW, Kim HS, Ihm HB. 1998a. Studies on the vegetation at the wetland of Suncheon-Man. *Bull Inst Litt Environ Mokpo Natl Univ* 15: 1-8.
- Ihm BS, Lee JS, Kim JW, Kim HS, Ihm HB. 1998b. Studies on the vegetation distribution and biomass at the wetland of Hampyung-Man. *Bull Ins Litt Environ Mokpo Natl Univ* 15: 9-20.
- Kalisz S. 1990. Soil seed banks. *Ecology* 71: 1226-1227.
- Kim CH, Lee KB, Kim JD, Cho TD, Lim MS. 2005. The study on the flora and vegetation of salt marshes of Dongjin-river estuary in Jeonbuk. *J Environ Sci* 14: 817-825.
- Koh JH. 2007. A study on the potential contribution of soil seed bank to the revegetation. *J Korean Environ Res Reveg Technol* 10: 99-109.
- Lee JS. 1989. On establishment of halophytes along tidal level gradient at salt marshes of Mankyong and Donjin river estuaries. PhD Dissertation. Seoul National University, Seoul, Korea. (in Korean with English abstract)
- Lee JS, Ihm BS, Myeong HH, Park JW, Kim HS. 2009. Soil environment analysis and habitat of halophyte for restoration in the salt marshes of southern and western coasts of Korea. *Korean J Plant Res* 22: 102-110.
- Lee JS, Kim JK. 1988. Factors affecting plant distribution in salt marsh of Mankyong river and Donjin river estuaries. *Nat Sci Res Kunsan Natl Univ* 3: 45-59.
- Lee SM. 2007. Secondary succession and seed bank structure of abandoned rice paddy field, and its restoration. Ms Thesis. Seoul Women's University, Seoul, Korea.
- Lee SM, Cho YC, Shin HC, Oh WS, Seol ES, Park SA, Lee CS. 2008. Successional changes in seed banks in abandoned rice fields im Gwangneung, central Korea. *J Ecol Field Biol.* 31: 269-276.
- Magurran AE. 2004. Measuring Biological Diversity. Blackwell, New York.
- McCune B, Mefford MJ. 1999. PC-ORD, Multivariate Analysis of Ecological Data. Version 4. MjM Software Design, Glenden Beach, OR.
- Middleton B. 1999. Wetland Restoration, Flood Pulsing, and Disturbance Dynamics. John Wiley & Sons, Inc., New York.
- Ministry of Environment. 2006. Development of Ecological Conservation and Management Techniques of Halophyte Communities on Salt Marshes of Southwestern Coasts in Korea. Korea Institute of Environmental Science and Technology, Seoul.
- Morton RD, Law R. 1997. Regional species pools and the assembly of local ecological communities. *J Theor Biol* 187: 321-331.
- National Research Council. 1992. Restoration of Aquatic Ecosystems: Science, Technology, and Public Policy. National Academy Press, Washington, DC.
- Nottage A, Robertson P. 2005. The Saltmarsh Creation Handbook: A Project Manager's Guide to the Creation of Saltmarsh and Intertidal Mudflat. The Royal Society for the Protection of Birds, Sandy, Bedfordshire.
- Pielou EC. 1966. The measurement of diversity in different types of biological collections. *J Theor Biol* 13: 131-144.
- Richter R, Stromberg JC. 2005. Soil seed banks of two montane riparian areas: implications for restoration. *Biodivers Conserv* 14: 993-1016.
- Roberts HA. 1981. Seed banks in soils. *Adv Appl Biol* 6: 1-55.
- Roovers P, Bossuyt B, Igodt B, Hermy M. 2006. May seed banks contribute to vegetation restoration on paths in temperate deciduous forest? *Plant Ecol* 187: 25-38.
- Simpson RL, Leck MA, Parker VT. 1989. Seed banks: general concepts and methodological issues. In: Ecology of Soil Seed Banks (Leck MA, Parker VT, Simpson RL, eds). Academic Press, London, pp 1-8.
- Society for Ecological Restoration International Science & Policy Working Group. 2004. The SER International Primer on Ecological Restoration. Society for Ecological

- Restoration International, Tucson, AZ.
- Stanturf JA, Schoenholtz SH, Schweitzer CJ, Shepard JP. 2001. Achieving restoration success: myths in bottomland hardwood forests. *Restor Ecol* 9: 189-200.
- Thompson K, Bakker JP, Bekker RM. 1997. *The Soil Seed Banks of North West Europe: Methodology, Density and Longevity*. Cambridge University Press, Cambridge.
- Toth LA. 1996. Restoring the hydrogeomorphology of the channelized Kissimmee River. In: *River Channel Restoration: Guiding Principles for Sustainable Projects* (Brookes A, Shields FD Jr, eds). John Wiley & Sons, Chichester, pp 369-383.
- Turner RE, Mendelssohn IA, McKee KL, Costanza R, Neill C, Sikora JP, Sikora WB, Swenson E. 1988. Wetlands hydrology and vegetation dynamics. *Proceedings of the National Wetland Symposium: mitigation of impacts and losses*, 1986 Oct 8-10, New Orleans, LA. Association of State Wetland Managers Berne, New York, pp 135-141.
- Wall DH, Palmer MA, Snelgrove PVR. 2001. Biodiversity in critical transition zones between terrestrial, freshwater, and marine soils sediments: processes, linkages, and management implications. *Ecosystems* 4: 418-420.
- Wang CH, Tang L, Fei SF, Wang JQ, Gao Y, Wang Q, Chen JK, Li B. 2009. Determinants of seed bank dynamics of two dominant helophytes in a tidal salt marsh. *Ecol Eng* 35: 800-809.
- White E, Tucker N, Meyers NM, Wilson J. 2004. Seed dispersal to revegetated isolated rainforest patches in North Queensland. *For Ecol Manage* 192: 409-426.
- Willems JH, Bik LPM. 1998. Restoration of high species density in calcareous grassland: the role of seed rain and soil seed bank. *Appl Veg Sci* 1: 91-100.
- Willson MF, Traveset A. 2000. The ecology of seed dispersal. In: *Seeds: The Ecology of Regeneration in Plant Communities* (Fenner M, ed). 2nd ed. CABI Publishing, Wallingford, pp 85-110.
- Wolters M, Bakker JP. 2002. Soil seed bank and driftline composition along a successional gradient on a temperate salt marsh. *Appl Veg Sci* 5: 55-62.
- Yim YJ, Kira T. 1975. Distribution of forest vegetation and climate in the Korean Peninsula I. Distribution of some indices of thermal climate. *Jpn J Ecol* 25: 77-88.
- Zhan X, Li L, Cheng W. 2007. Restoration of *Stipa krylovii* steppes in inner Mongolia of China: assessment of seed banks and vegetation composition. *J Arid Environ* 68: 298-307.
- Zobel M. 1997. The relative role of species pools in determining plant species richness: an alternative explanation of species coexistence? *Trends Ecol Evol* 12: 266-269.
- Zobel M, van der Maarel E, Dupré C. 1998. Species pool: the concept, its determination and significance for community restoration. *Appl Veg Sci* 1: 55-66.