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# The effect of phosphorus removal from sewage on the plankton community in a hypertrophic reservoir

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## Abstract

**Background:** When developing water quality improvement strategies for eutrophic lakes, questions may arise about the relative importance of point sources and nonpoint sources of phosphorus. For example, there is some skepticism regarding the effectiveness of partial reductions in phosphorus loading; because phosphorus concentrations are too high in hypertrophic lakes, in-lake phosphorus concentrations might still remain within typical range for eutrophic lakes even after the reduction of phosphorus loading. For this study, water quality and the phytoplankton and zooplankton communities were monitored in a hypertrophic reservoir (Lake Wangsong) before and after the reduction of phosphorus loading from a point source (a sewage treatment plant) by the installation of a chemical phosphorus-removal process.

**Results:** Before phosphorus removal, Lake Wangsong was classified as hypertrophic with a median phosphorus concentration of 0.232 mg L<sup>-1</sup> and a median chlorophyll-a concentration of 112 mg L<sup>-1</sup>. The dominant phytoplankton were filamentous cyanobacteria for the most of the ice-free season. Following the installation of the advanced treatment process, phosphorus concentrations were reduced to 81 mg L<sup>-1</sup>, and the N/P atomic ratio increased from 42 to 102. Chlorophyll-a concentrations decreased to 42 µg L<sup>-1</sup>, and the duration of cyanobacterial dominance was confined to the summer season. Cyanobacteria in spring and autumn were replaced by diatoms and cryptomonads. Filamentous cyanobacteria in summer were replaced by colony-forming unicellular *Microcystis* spp. It was remarkable that zooplankton biomass increased despite the decrease in phytoplankton biomass, and especially cladoceran zooplankton which increased drastically. These responses to the reduction of point source P loading to Lake Wangsong imply that reducing the point source P loading can have a big impact even when nonpoint sources account for a large fraction of the total annual phosphorus loading.

**Conclusions:** Our results also show that the phytoplankton community can shift to decreased cyanobacterial dominance and the zooplankton community can shift to higher cladoceran dominance, even when phosphorus concentrations remain within the typical range for eutrophic lakes following the reduction of phosphorus loading.

**Keywords:** Eutrophication, Phosphorus removal, Phytoplankton, Point source, Zooplankton

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## Background

Eutrophication is the most common water quality problem in lakes and reservoirs (Azevedo et al. 2015; Reed-Andersen et al. 2000), and phosphorus is the major limiting factor of eutrophication (Schindler 1977; Withers and Jarvie 2008). Phosphorus (P) comes mostly from animal excretion and fertilizer, and the sources are classified as point sources or nonpoint sources (Kundu et al. 2015; Carson et al. 2015). In most rural watersheds, nonpoint P sources such as agricultural fields and forest are the major sources of the total P load. However, if there is significant urban development within the watershed, sewage discharge can contribute a considerable portion of P loading (Neal et al. 2000; Jarvie et al. 2006). In watersheds consisting of complex terrain with urban and rural land use, P loading from agricultural fields is usually larger than the P loading from sewage. However, the export of P from nonpoint sources is exported mostly on rainy days in the form of storm runoffs (Li et al. 2015). On the contrary, P loading from point sources have uniform flow rates, and on dry days, their contribution can be larger than those of nonpoint sources.

Hydraulic residence time can be another factor that determines the relative importance of storm runoff in lakes and reservoirs (Dillon 1975; Brett and Benjamin 2008). Shock loading from agricultural nonpoint sources can be very large during a storm event, but a large portion of the P loading can be flushed out of reservoirs with short residence times. Therefore, there are still many important questions remaining concerning the relative importance of point and nonpoint P loading. There might be some skepticism regarding the effectiveness of reductions in point source P loading which result in only a partial reduction of the annual P loading (Mainstone and Parr 2002).

Point sources of P include raw sewage discharge or effluent from sewage treatment plants which do not have chemical phosphorus-removal processes. The sewage treatment process can be divided into three phases: primary treatment, secondary treatment, and tertiary treatment (or advanced treatment). Commonly, in developed countries, sewage is treated using secondary treatment processes, biological treatment processes designed mainly for biological oxygen demand (BOD) removal (Cullen and Forsberg 1988). Because P concentrations in the effluent from secondary treatment (1–2 mg P L<sup>-1</sup>) is 30–60 times higher than the typical criterion for eutrophication (typically 0.03 mg P L<sup>-1</sup>), advanced P treatment using chemical precipitation has been incorporated in some sewage treatment plants in Korea that discharge effluent into sensitive surface waters. The legal standards for P concentrations in effluent from treatment plants are between 0.2 and 0.5 mg P L<sup>-1</sup>, still substantially higher than the criterion for eutrophication.

In hypertrophic lakes where P concentrations are too high even after a reduction of point source P loading, the in-lake P concentrations may still remain in the typical range for eutrophic lakes (Schindler 2006). As part of efforts to provide a scientific basis for decision-making and developing strategies for water quality management, detailed information on the consequences of a reduction in P loading and especially the effects on water quality and the phytoplankton and zooplankton communities is required.

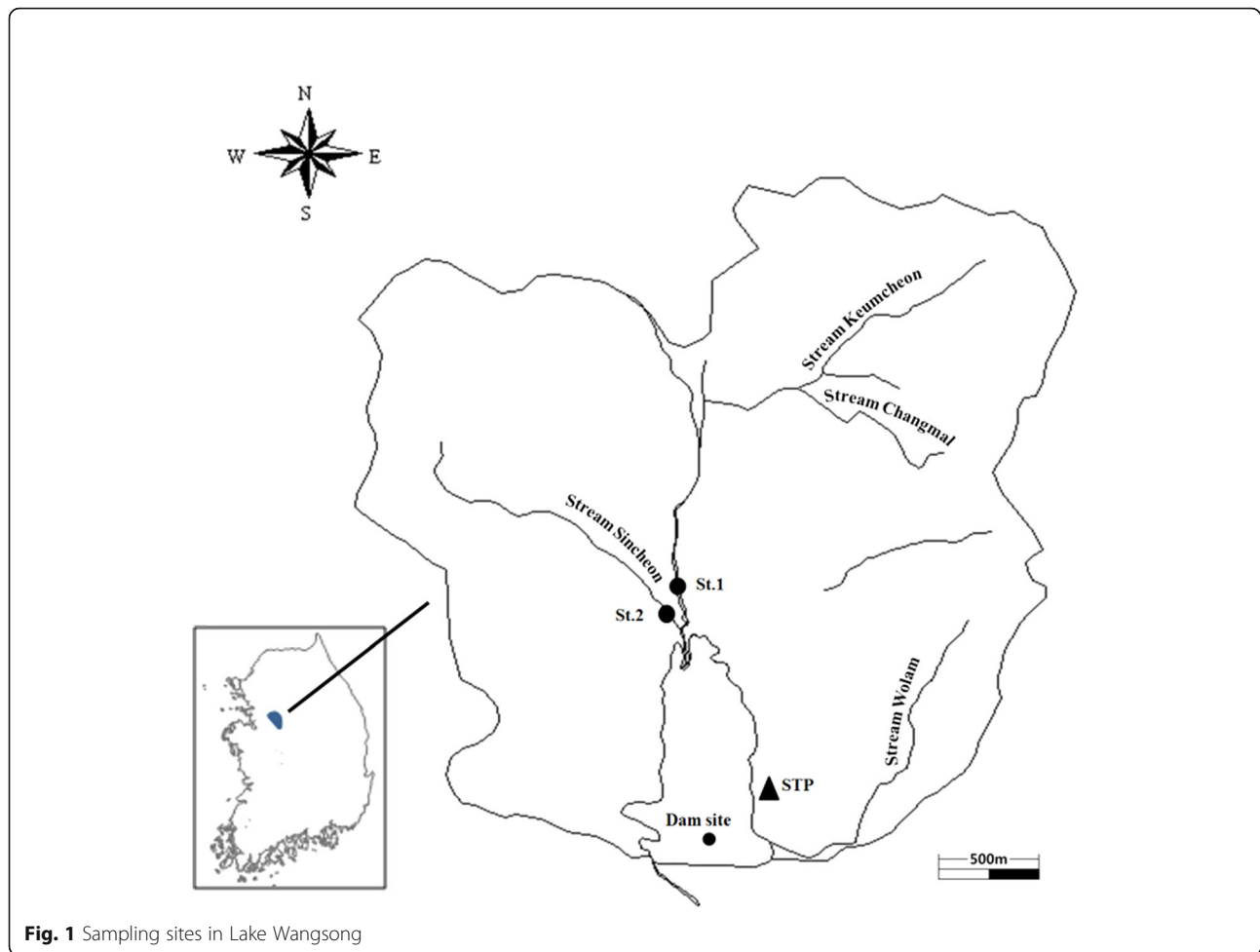
Lake Wangsong is a hypertrophic reservoir located in a watershed with mixed urban and rural land use (Table 1). The lake has been hypertrophic since a sewage treatment plant (STP) was constructed on the shore of the lake. Because the STP effluent contained high P concentrations, the discharge from the STP was identified as the major cause of eutrophication, and the advanced treatment process was added to the plant to reduce P concentration in the STP effluent. In this study, water quality and the phytoplankton and zooplankton communities were surveyed before and after the installation of the advanced sewage treatment process in order to examine the effectiveness of the P removal process for the water quality improvement in a hypertrophic lake. We also examined the effect of the reduction in P loading on the phytoplankton and zooplankton communities in this lake.

## Methods

Lake Wangsong was constructed on the Hwangkujichon Stream (Uiwang city) for providing irrigation water for agriculture (Fig. 1). Rapid urbanization within the watershed resulted in an increase in sewage discharge, thus discharging a large amount of P into the lake. In 1999, the STP was constructed at the shore of the lake, and the advanced treatment process using chemical precipitation was added in August of 2007 to reduce P concentrations in the effluent.

**Table 1** Morphological characteristics of Lake Wangsong

Watershed area (km <sup>2</sup> )	15.5
Mean depth (m)	2.5
Total water storage capacity (10 <sup>3</sup> m <sup>3</sup> )	2079
Maximum surface area (km <sup>2</sup> )	0.86
Hydraulic residence time (days)	49
Land use (ha)	
Agriculture (upland field)	290 (18.6 %)
Agriculture (paddy field)	320 (20.6 %)
Forest	685 (44.1 %)
Residential area	182 (11.7 %)
Others	78 (5.0 %)
Total population	27,994
Population density (pop. km <sup>-2</sup> )	1806



**Fig. 1** Sampling sites in Lake Wangsong

The phytoplankton and zooplankton communities and water quality were surveyed before the advanced treatment operation (2003 and 2004) and after the advanced treatment (2008 and 2009). Water samples were collected from the center of the lake at 0, 2, and 5-m depths. Two main inflowing streams were surveyed in order to assess P loading from the watershed in the dry season. In this study, water quality measurements of the STP effluent were measured and, also, the data from the STP management office were employed together for calculating the P loading from the STP effluent.

Water samples were collected using an inflatable boat and a horizontal Van Dorn water sampler. The water samples were transported in a cooler and stored in a refrigerator until analysis. Dissolved oxygen was measured in situ with a DO meter (YSI, USA). All water quality measurements were conducted according to standard methods (American Public Health Association 2005) except chemical oxygen demand (COD). Total phosphorus (TP) concentration was measured using the ascorbic acid method following persulfate digestion. Chlorophyll-a concentration was measured by the

trichromatic spectrophotometric method. Total nitrogen (TN) was measured by the cadmium reduction method following persulfate digestion. Suspended solids (SS) was measured gravimetrically after filtration by GF/F filter. Biological oxygen demand (BOD) was measured by using a DO meter. COD measurements used the permanganate method, the official standard method in Korea (MOE 2009).

Water quality data collected at the surface of the lake by the Gyeonggido local province was also employed in addition to the data collected by this study in order to increase statistical significance in the comparison of water quality before and after the advanced treatment. Stream discharge of inflowing streams was measured using a magnetic flow velocity meter and the current cross-section method. P loading from nonpoint sources in the watershed was estimated by multiplying the unit export coefficients of phosphorus by the area of each land use type as suggested by the Korean Ministry of the Environment (MOE 2014).

Phytoplankton samples were collected in 500-mL polyethylene bottles at 0.5-m depth and preserved with

Lugol's solution. Cell densities were measured using a Sedgewick-Rafter counting chamber and an X300 microscope (Olympus BX50). Zooplankton samples were collected with a plankton net (63- $\mu\text{m}$  mesh) using a slow vertical tow from the bottom (typically 5–6-m depth) to the surface of the lake. Zooplankton samples were preserved with 4 % sucrose formalin (Steedman 1976). The volume of water filtered by a zooplankton net was calculated by multiplying the aperture area of the net by the towing distance, assuming there is no significant loss of filtering efficiency through a 5-m towing.

**Results**

**Nutrient loading and water quality**

Water quality parameters in the two main inflowing streams and the STP effluent are shown in Table 2. In general, the nutrient status in the inflowing streams indicates eutrophic conditions. The median TP concentration was 0.131 and 0.190 mg P L<sup>-1</sup> at St.1 and St.2, respectively. TP concentrations in the STP effluent were much higher than either St.1 or St.2 in 2003–2007 (median TP 0.633 mg P L<sup>-1</sup>). Following the phosphorus removal (in 2008), TP concentrations in the effluent were lower (median 0.310 mg P L<sup>-1</sup>).

The annual average daily P loading from nonpoint sources in the Lake Wangsong watershed was 7.5 kg P day<sup>-1</sup>, which was higher than the P loading from the major point source (STP effluent), 3.7 kg P day<sup>-1</sup> (Gyeonggi Research Institute 2011). But most of the annual P loading from agricultural nonpoint sources is concentrated in a few rain events during the summer monsoon season in Korea, and only limited amounts of P export occur on dry days. When the daily P loading was measured in dry seasons, the P discharge from the STP effluent accounted for a larger portion (56 and 60 %) than the P loading from the two main inflowing streams (Table 3).

In August of 2007, the advanced P removal treatment started operation. With the reduction of P loading, TP

**Table 3** Comparison of daily phosphorus loading on dry days from the watershed (St.1 and St.2) and sewage treatment plant effluent (STP)

Date		St.1	St.2	STP
25 Sep. 2008	Flow rate (m <sup>3</sup> day <sup>-1</sup> )	15,314	7733	11,584
	P loading (kg day <sup>-1</sup> )	1.75 (26.3 %)	0.91 (13.7 %)	3.98 (60.0 %)
31 Oct. 2008	Flow rate (m <sup>3</sup> day <sup>-1</sup> )	10,152	6229	11,910
	P loading (kg day <sup>-1</sup> )	1.71 (28.7 %)	0.9 (15.1 %)	3.37 (56.2 %)

concentrations in the surface have decreased from 0.232 to 0.081 mg P L<sup>-1</sup> (median), a reduction of 65 % (Table 2). But even after the advanced treatment, TP concentrations in the STP effluent were ten times higher than the typical criterion for eutrophic conditions (0.03 mg P L<sup>-1</sup>). Consequently, TP in Lake Wangsong also exceeded the threshold of eutrophication by a factor of 2.5.

Because the advanced treatment was focused on the chemical removal of P, which has a lower removal efficiency than nitrogen, TN concentrations did not decrease as much as TP concentrations. With the higher removal efficiency of P than N, the atomic N/P ratio in the lake increased from 68 to 154 following the start of the advanced treatment. Thus, the atomic N/P was much higher than the Redfield ratio (16) (Redfield 1958) implying P limitation of algal growth in Lake Wangsong.

Suspended solids (SS) in the lake did not change after the advanced treatment. The median SS was 19 mg L<sup>-1</sup>, suggesting that the seston is composed of mostly inorganic particles, possibly arising from the bottom sediment in this shallow reservoir. Assuming that the chlorophyll-a content of algal cells is commonly 1 % of dry weight and the median chlorophyll-a concentration of 42.2 ug L m<sup>-1</sup> in Lake Wangsong, algal cells would account for an algal biomass of

**Table 2** Water quality of inflowing streams and lake surface (median, mg L<sup>-1</sup>; Chl.a, ug L<sup>-1</sup>, (25th–75th percentile))

	BOD	COD	SS	TN	TP	Chl.a	BOD/COD	N/P (atomic)	Chl.a/TP	Number
Inflow St.1 (2008–2009)	4.6 (3.1–5.0)	6.4 (5.9–6.4)	45.7 (32.0–71.5)	3.73 (3.71–4.89)	0.190 (0.124–0.244)		0.16 (0.09–0.23)	51 (38–54)		10
Inflow St.2 (2008–2009)	2.7 (1.3–5.0)	6.4 (5.0–8.1)	11.9 (5.7–44.1)	3.76 (3.30–4.37)	0.131 (0.113–0.222)		0.42 (0.23–0.79)	48 (38–68)		10
STP effluent (2003–2007)	10.4 (8.3–12.7)	14.9 (12.3–17.7)	12.5 (7.1–23.3)	7.9 (5.9–13.6)	0.633 (0.270–1.299)		0.65 (0.50–0.77)	28 (21–44)		24
STP effluent (2008–2009)	1.6 (1.0–2.2)	6.8 (6.0–7.8)	1.9 (1.4–3.6)	10.6 (8.9–12.5)	0.310 (0.185–0.355)		0.25 (0.15–0.32)	78 (61–88)		29
Lake surface (May 1999–Jun. 2007)	11.4 (8.7–13.7)	15.8 (13.2–18.4)	20.4 (16.0–30.7)	4.75 (3.47–5.87)	0.232 (0.181–0.284)	112.4 (53.9–163.7)	0.70 (0.51–0.87)	42 (27–70)	0.54 (0.22–0.84)	42
Lake surface (Aug. 2007–Nov. 2009)	5.2 (4.1–7.1)	10.2 (8.7–11.3)	19.1 (13.9–30.8)	3.56 (2.31–4.93)	0.081 (0.054–0.137)	42.2 (24.9–110.8)	0.59 (0.47–0.69)	102 (68–154)	0.61 (0.28–1.15)	20

approximately 4 mg L<sup>-1</sup>, much lower than the SS. Therefore, inorganic particles would account for 15 mg L<sup>-1</sup> in the SS. In Lake Wangsong, resuspension of sediment is likely common on windy days. The decrease in BOD following the reduction in P loading affected the hypolimnetic DO concentrations, a common criterion for eutrophication (Horne and Goldman 1994). Anoxic conditions in the hypolimnion clearly developed below 4-m depth in September 2003 and May to August 2004, whereas the hypolimnion was oxic in 2009.

#### Phytoplankton and zooplankton

With decreased in-lake TP, chlorophyll-a concentrations decreased drastically from a median of 112.4 mg m<sup>-3</sup> before the advanced treatment to 42.2 mg m<sup>-3</sup> after the advanced treatment; the decrease in chlorophyll-a was 62 %, similar to the decrease in TP (65 %, Table 2). The maximum cell density for cyanobacteria also decreased drastically from 25 × 10<sup>5</sup> cells mL<sup>-1</sup> to 9 × 10<sup>5</sup> cells mL<sup>-1</sup> (Table 4). The phytoplankton species composition changed together with a decrease in algal standing crop. In 2003 and 2004, cyanobacteria were the dominant phytoplankton species during most of the ice-

free season from April to December. It was remarkable that cyanobacteria were dominant even in winter. By contrast, after the advanced treatment, cyanobacteria were dominant in only 4 samples out of 12 monthly samples (Table 4). In the summer months, cyanobacteria were still dominant, but in spring and autumn, cryptomonads replaced cyanobacteria. In the winter samples, diatoms were the dominant phytoplankton. In a statistical comparison, the density of diatoms and cyanobacteria showed significant difference between before and after the advanced treatment ( $p < 0.05$ ).

With the change in phytoplankton species composition, the dominant zooplankton also changed (Table 5). The most remarkable change was that large-sized cladoceran zooplankton (*Daphnia galeata*) reached the maximum density in spring and autumn when cyanobacteria were not dominant. Before the advanced treatment started, the maximum cladoceran density was 250 ind. L<sup>-1</sup> in July and August, whereas in April 2008, the density of cladocerans was 1242 ind. L<sup>-1</sup>. In the statistical analysis, the density and dominance of cladocerans between before and after the advanced treatment showed significant differences.

**Table 4** Cell densities of three major phytoplankton taxa and dominant species before and after advanced sewage treatment

Year	Month	Cell density (cells mL <sup>-1</sup> )					Dominant species (by biomass, µgC L <sup>-1</sup> )
		Bacillariophyceae	Cyanophyceae	Chlorophyceae	Others	Total	
2003	Sep.	6408	497,304	534	971	505,217	<i>Oscillatoria</i> sp.
	Nov.	4269	213,537	5817	123	223,746	<i>Microcystis</i> sp.
	Dec.	115	59,486	3326	1840	64,767	<i>Oscillatoria</i> sp.
2004	Mar.	381	757	545	25	1708	<i>Closterium</i> sp.
	Apr.	185	130,010	122,333	3933	256,461	<i>Anabaena</i> sp.
	May	43	1,474,200	54,871	1207	1,530,321	<i>Lyngbya</i> sp.
	Jun.	58	2,515,477	1234	13,247	2,530,016	<i>Microcystis</i> sp.
	Jul.	3385	508,721	1362	34	513,502	<i>Lyngbya</i> sp.
	Aug.	299	1,378,199	59	595	1,379,152	<i>Microcystis</i> sp.
	Advanced treatment start						
2007	Nov.	36	7088	89	1148	8361	<i>Cryptomonas ovata</i>
	Dec.	2196	673	414	61	3344	<i>Stephanodiscus hantzschii</i>
2008	Apr.			1055	1570	2625	<i>Cryptomonas ovata</i>
	May	2	796	39,744	389	40,931	<i>Pediastrum duplex</i>
	Jun.	2	897,039	276	22	897,339	<i>Microcystis aeruginosa</i>
	Jul.	230	33,155	469	533	34,386	<i>Microcystis aeruginosa</i>
	Aug.	20	160,053	111	2215	162,399	<i>Microcystis wesenbergii</i>
	Sep.	210	30,197	2806	393	33,606	<i>Oscillatoria sancta</i>
	Dec.	5581	13	1148	47	6789	<i>Stephanodiscus hantzschii</i>
	2009	Mar.	24		190	315	528
May		8		3524	1168	4701	<i>Cryptomonas</i> sp.
Sep.		12	2349	24	133	2518	<i>Cryptomonas</i> sp.

**Table 5** Standing crop and dominant zooplankton species

Year	Month	Standing crop (ind. L <sup>-1</sup> )				Dominant species
		Cladocera	Copepoda	Rotifera	Total	
2003	Sep.	57	19	13	89	<i>Bosmina longirostris</i>
	Nov.	2	18	21	41	<i>Cyclops vicinus</i>
	Dec.	21	14	122	157	<i>Asplanchna priodonta</i>
2004	Mar.	1	13	85	98	<i>Cyclops vicinus</i>
	Apr.	44	29	2731	2804	<i>Cyclops vicinus</i>
	May	0	1	10	11	Nauplius
	Jun.	48	298	1805	2151	<i>Thermocyclops taihokuensis</i>
	Jul.	254	368	1555	2177	<i>Thermocyclops taihokuensis</i>
	Aug.	243	271	989	1504	Nauplius
Advanced treatment start						
2007	Nov.	5	18	20	43	<i>Bosmina longirostris</i>
	Dec.	12	58	12	88	<i>Cyclops vicinus</i>
2008	Apr.	1242	1089	1531	3862	<i>Daphnia galeata</i>
	May	550	342	626	1518	<i>Cyclops vicinus</i>
	Jun.	49	162	54	266	<i>Daphnia galeata</i>
	Jul.	77	212	1384	1674	Copepodite
	Aug.	22	235	1561	1818	<i>Brachionus forficula</i>
	Sep.	97	306	2704	3105	<i>Keratella cochlearis</i>
	Oct.	510	1729	1758	3997	<i>Daphnia galeata</i>
	Dec.	217	466	3616	4299	<i>Cyclops vicinus</i>
2009	Mar.	21	120	154	295	<i>Cyclops vicinus</i>
	May	0	0	2275	2275	<i>Daphnia galeata</i>
	Sep.	9	82	493	574	<i>Cyclops vicinus</i>

## Discussion

The annual P loading to lakes from point sources and nonpoint sources is commonly comprised mostly of agricultural nonpoint P sources in watersheds with mixed land uses. However, in reservoirs with short hydraulic residence time, storm runoff from nonpoint sources is not stored for a long time within the reservoir, especially in the rainy season. Because Korea is located in the summer monsoon region, most of the annual rainfall occurs in summer. Short hydraulic residence times in reservoirs can be a critical factor affecting nutrient concentrations in reservoirs. In the dry season, the relative importance of point sources is larger than nonpoint sources, because nonpoint sources do not export nutrients during periods with minimum stream flow. In this study, the importance of P removal in the STP effluent was obviously manifested in the water quality improvement after the advanced treatment.

Even after the advanced treatment, Lake Wangsong remained eutrophic. TP and chlorophyll-a concentrations were higher than the criteria for eutrophic lakes (Wetzel 2001), mainly because TP in the STP effluent was still much higher than the typical criterion for

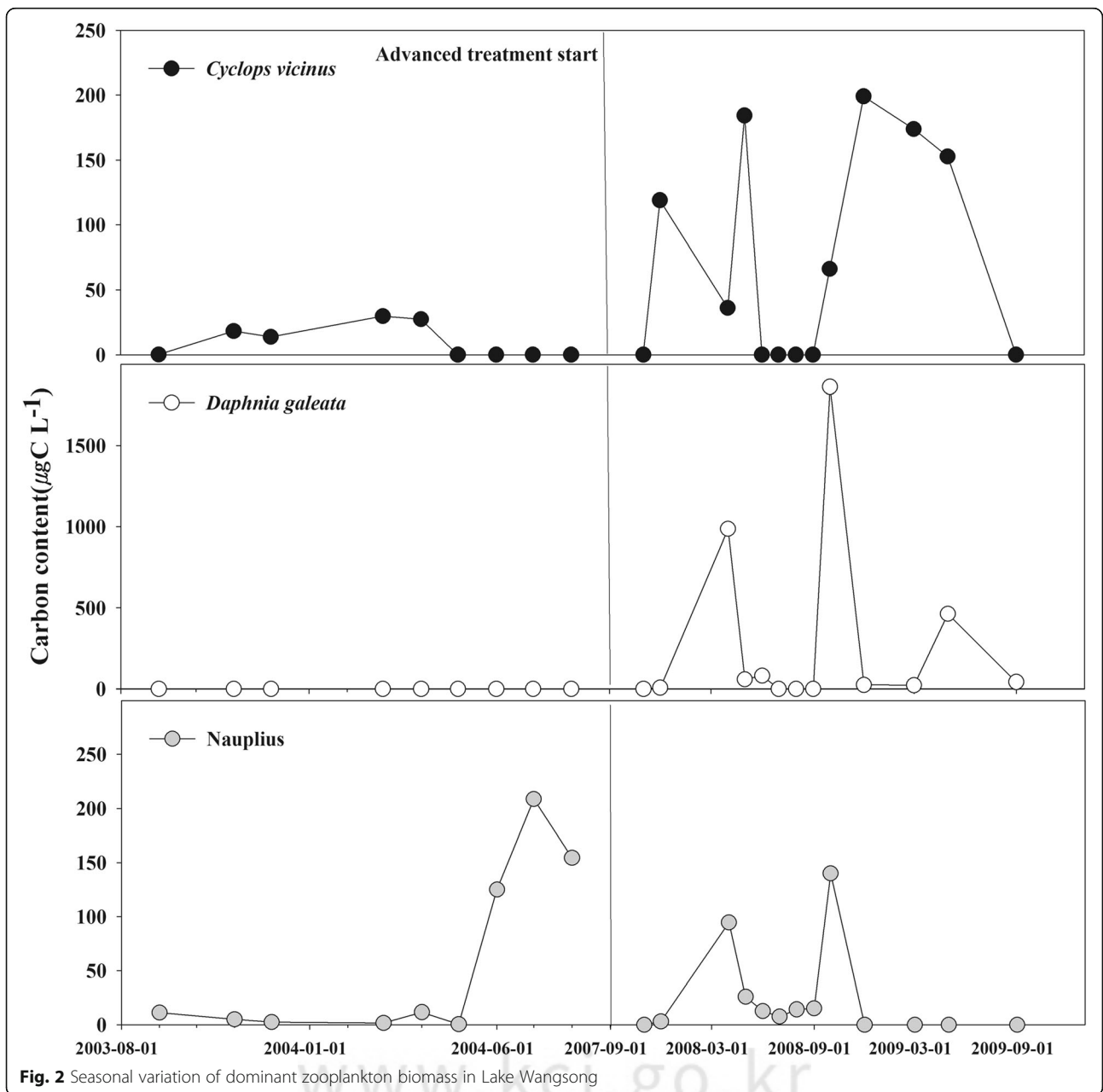
eutrophication. TP can be reduced to as low as 0.01 mg P L<sup>-1</sup> in STP effluent with advanced chemical treatment. In Korea, the phosphorus standards for STP effluent are in the range of 0.2 to 2.0 mg L<sup>-1</sup>, still much higher than the eutrophication criterion. Therefore, even if sewage is treated according to the government standard for P concentration in the effluent, sewage can be a main cause of eutrophication and further reductions in P in STP effluent are strongly needed for the control of eutrophication.

Importantly, we observed that the dominance of cyanobacteria decreased with P concentrations, even though P concentrations remained in the eutrophic level. The N/P ratio increased due to the P removal in the STP effluent, which might have provided favorable conditions for algae other than cyanobacteria. A low N/P ratio is known to be favorable for cyanobacteria, because N can be a temporary limiting factor and N-fixing cyanobacteria can take advantage of this (Gu and Alexander 1993). Temporary nitrogen depletion during algal blooms can be a controlling factor in the competition among phytoplankton species; that is, decreases in P can provide favorable conditions for other

algal species and inhibit the dominance of filamentous cyanobacteria (Fulton 1988). Before the start of the advanced treatment, most of the dominant phytoplankton were filamentous cyanobacteria with many species having the potential to carry out N-fixation, whereas filamentous cyanobacteria were dominant only in 1 sample out of 12 monthly samples after the advanced treatment.

The change of the zooplankton community in Lake Wangsong was also obvious. In 2003 and 2004, rotifers were the dominant zooplankton, whereas cladoceran species were the dominant zooplankton in 2008 and 2009 (Table 5). A remarkable result was that zooplankton biomass increased after the advanced treatment even

though the phytoplankton biomass decreased. Generally, zooplankton standing crop is associated with the availability of phytoplankton, the so-called bottom-up effect (Sinistro 2009). In addition, copepod and cladoceran densities increased drastically and rotifer density increased slightly. The increase in cladoceran density is usually regarded as a positive change for water quality due to their large filtering capacity which can result in an increase in water clarity (Sommer et al. 2001; Schrage and Downing 2004). Especially, the maximum cladoceran densities in April 2008 and October 2009 may have caused clear water phases due to the high density of *D. galeata* (Fig. 2). In 2003 and 2004, the



dominant phytoplankton were filamentous cyanobacteria during all seasons, and these species are generally known as inedible prey for zooplankton (Krevš et al. 2010). The change in zooplankton species composition can be explained as the result of change of phytoplankton community from inedible filamentous cyanobacteria to edible diatoms and cryptomonads which were dominant in spring and autumn of 2009 (Bomi et al. 2013).

## Conclusions

The reduction of phosphorus from STP effluent resulted in a significant reduction of in-lake P concentrations in Lake Wangsong, which in turn effected a shift in the phytoplankton community. Even though P concentrations remained within the typical range for eutrophic conditions following the reduction in P loading, all the indicators of water quality and aquatic ecosystem health showed improvements: decreased phytoplankton density, decreased hypoxia in the hypolimnion, a shift from cyanobacteria to diatoms in cold seasons, a shift from filamentous cyanobacteria to colony-forming unicellular cyanobacteria, and increased cladoceran zooplankton populations which can improve water clarity and facilitate the transfer of energy through the grazing food chain. This implies that P removal from STP effluent can be important for improving water quality in hypertrophic reservoirs even if the annual P loading from nonpoint sources is larger than the P loading from sewage effluent.

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## Availability of data and materials

Please contact the author for data requests.

## Authors' contributions

All authors contributed extensively to the work presented in this paper. BK and SJ designed the study and wrote the main paper. YL and JL collected the zooplankton data and analyzed the results. SJ, JK, and YC collected and analyzed the phytoplankton data and conducted the data analysis. KK, JO, and AR analyzed the water quality data. All authors discussed the results and implications and commented on the manuscript at all stages.

## Competing interests

The authors declare that they have no competing interests.

## Consent for publication

Not applicable.

## Ethics approval and consent to participate

Not applicable.

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